



Designation: E1525 – 02 (Reapproved 2023)

# Standard Guide for Designing Biological Tests with Sediments<sup>1</sup>

This standard is issued under the fixed designation E1525; the number immediately following the designation indicates the year of original adoption or, in the case of revision, the year of last revision. A number in parentheses indicates the year of last reapproval. A superscript epsilon ( $\epsilon$ ) indicates an editorial change since the last revision or reapproval.

## 1. Scope\*

1.1 As the contamination of freshwater and saltwater ecosystems continues to be reduced through the implementation of regulations governing both point and non-point source discharges, there is a growing emphasis and concern regarding historical inputs and their influence on water and sediment quality. Many locations in urban areas exhibit significant sediment contamination, which poses a continual and long-term threat to the functional condition of benthic communities and other species inhabiting these areas (1).<sup>2</sup> Benthic communities are an important component of many ecosystems and alterations of these communities may affect water-column and nonaquatic species.

1.2 Biological tests with sediments are an efficient means for evaluating sediment contamination because they provide information complementary to chemical characterizations and ecological surveys (2). Acute sediment toxicity tests can be used as screening tools in the early phase of an assessment hierarchy that ultimately could include chemical measurements or bioaccumulation and chronic toxicity tests. Sediment tests have been applied in both saltwater and freshwater environments (2-6). Sediment tests have been used for dredge material permitting, site ranking for remediation, recovery studies following management actions, and trend monitoring. A particularly important application is for establishing contaminant-specific effects and the processes controlling contaminant bioavailability(7).

1.3 This guide is arranged as follows:

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<sup>1</sup> This guide is under the jurisdiction of ASTM Committee E50 on Environmental Assessment, Risk Management and Corrective Action and is the direct responsibility of Subcommittee E50.47 on Biological Effects and Environmental Fate.

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<sup>2</sup> The boldface numbers in parentheses refer to the list of references at the end of this standard.

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1.4 The values stated in SI units are to be regarded as the standard. The values given in parentheses are for information only.

1.5 *This standard does not purport to address all of the safety concerns, if any, associated with its use. It is the responsibility of the user of this standard to establish appropriate safety, health, and environmental practices and determine the applicability of regulatory limitations prior to use.* For specific hazard statements, see Section 7.

1.6 *This international standard was developed in accordance with internationally recognized principles on standardization established in the Decision on Principles for the Development of International Standards, Guides and Recommendations issued by the World Trade Organization Technical Barriers to Trade (TBT) Committee.*

## 2. Referenced Documents

### 2.1 ASTM Standards:<sup>3</sup>

- D1129 Terminology Relating to Water
- D4447 Guide for Disposal of Laboratory Chemicals and Samples
- E724 Guide for Conducting Static Short-Term Chronic Toxicity Tests Starting with Embryos of Four Species of Saltwater Bivalve Molluscs
- E729 Guide for Conducting Acute Toxicity Tests on Test Materials with Fishes, Macroinvertebrates, and Amphibians
- E943 Terminology Relating to Biological Effects and Environmental Fate (Withdrawn 2023)<sup>4</sup>
- E1023 Guide for Assessing the Hazard of a Material to Aquatic Organisms and Their Uses
- E1367 Test Method for Measuring the Toxicity of Sediment-Associated Contaminants with Estuarine and Marine Invertebrates

<sup>3</sup> For referenced ASTM standards, visit the ASTM website, www.astm.org, or contact ASTM Customer Service at service@astm.org. For *Annual Book of ASTM Standards* volume information, refer to the standard's Document Summary page on the ASTM website.

<sup>4</sup> The last approved version of this historical standard is referenced on www.astm.org.

\*A Summary of Changes section appears at the end of this standard

E1383 Guide for Conducting Sediment Toxicity Tests with Freshwater Invertebrates (Withdrawn 1995)<sup>4</sup>

E1391 Guide for Collection, Storage, Characterization, and Manipulation of Sediments for Toxicological Testing and for Selection of Samplers Used to Collect Benthic Invertebrates

E1563 Guide for Conducting Short-Term Chronic Toxicity Tests with Echinoid Embryos

E1611 Guide for Conducting Sediment Toxicity Tests with Polychaetous Annelids

E1676 Guide for Conducting Laboratory Soil Toxicity or Bioaccumulation Tests with the Lumbricid Earthworm *Eisenia Fetida* and the Enchytraeid Potworm *Enchytraeus albidus*

E1688 Guide for Determination of the Bioaccumulation of Sediment-Associated Contaminants by Benthic Invertebrates

E1706 Test Method for Measuring the Toxicity of Sediment-Associated Contaminants with Freshwater Invertebrates

IEEE/ASTM SI-10 Standard for Use of the International System of Units (SI): The Modern Metric System

2.2 Other Standards:

Title 29 Code of Federal Regulations 1910.132 (f)<sup>5</sup>

### 3. Terminology

#### 3.1 Definitions:

3.1.1 The words “must,” “should,” “may,” “can,” and “might” have very specific meanings in this guide. “Must” is used to express an absolute requirement, that is, to state that the test ought to be designed to satisfy a specific condition, unless the purpose of the test requires a different design. “Must” is used only in connection with the factors that apply directly to the acceptability of the test. “Should” is used to state that the specified conditions are recommended and ought to be met in most tests. Although a violation of one “should” is rarely a serious matter, violation of several will often render the results questionable. Terms such as “is desirable,” “is often desirable,” and “might be desirable” are used in connection with less important factors. “May” is used to mean “is (are) allowed to,” “can” is used to mean “is (are) able to,” and “might” is used to mean “could possibly.” Thus, the classic distinction between “may” and “can” is preserved, and “might” is never used as a synonym of either “may” or “can.”

3.1.2 For definitions of terms used in this guide, refer to Guide E729, Terminologies D1129 and E943, and Guide E1023. For an explanation of the units and symbols, refer to IEEE/ASTM SI-10.

#### 3.2 Definitions of Terms Specific to This Standard:

3.2.1 *bioaccumulation*—the net uptake of a material by an organism from its environment through exposure by means of water and food.

3.2.2 *concentration*—the ratio of the weight or volume of test material(s) to the weight or volume of test sample.

3.2.3 *control sediment*—a sediment that is essentially free of contaminants and is used routinely to assess the acceptability of a test.

3.2.4 *elutriate*—the water and soluble portion extracted from the sediment.

3.2.5 *exposure*—contact with a chemical or physical agent.

3.2.6 *overlying water*—the water placed over the solid phase of a sediment in the test chamber for the conduct of the biological test; this may also include the water used to manipulate the sediments. In field situations, the water column above the sediment/water interface.

3.2.7 *pore water/interstitial water*—water occupying space between sediment or soil particles.

3.2.8 *reference sediment*—a whole sediment near the area of concern used to assess sediment conditions exclusive of material(s) of interest.

3.2.9 *sediment*—(1) particulate material that usually lies below water and (2) formulated particulate matter that is intended to lie below water in a test.

3.2.10 *spiked sediment*—a sediment to which a material has been added for experimental purposes.

3.2.11 *suspension*—a slurry of sediment and water.

3.2.12 *toxicity*—the property of a material or combination of materials to affect organisms adversely.

3.2.13 *whole sediment*—sediment and associated pore water that has had minimal manipulation following collection or formulation.

### 4. Application

4.1 An ASTM guide outlines a series of options or instructions and does not recommend a specific course of action. The purpose of a guide is to offer guidance, based on a consensus of viewpoints, but not to establish a fixed procedure. A guide is intended to increase the awareness of the user to available techniques in a given subject area and to provide information from which subsequent evaluation and standardization can be derived.

4.2 This guide provides general interpretative guidance on the selection, application, and interpretation of biological tests with sediments. As such, this guide serves as a preface to other ASTM documents describing methods for sediment collection, storage, and manipulation (Guide E1391); and toxicity or bioaccumulation tests with sediment ( Guides E724, E1367, E1391, E1611, E1563, E1688, and Test Method E1706). Much of the guidance presented in this standard is also applicable to toxicity testing of soils (Guide E1676). This guide serves as an introduction and summary of sediment testing and is not meant to provide specific guidance on test methods. Rather, its intent is to provide information necessary to accomplish the following:

4.2.1 Select a sediment exposure strategy appropriate to the assessment need. For example, a suspended phase exposure is relevant to the evaluation of dredged sediments for disposal at a dispersive aquatic site. (See Annex A1).

4.2.2 Select the test organism and biological endpoints appropriate to the desired exposure and aquatic resources at

<sup>5</sup> Available from Superintendent of Documents, U.S. Government Printing Office, Washington DC 20402.

risk. For example, the potential for water quality problems and subsequent effects on oyster beds may dictate the use of sediment elutriate exposures with bivalve larvae (Guide E724).

4.2.3 Establish an experimental design consistent with the objectives of the sediment evaluation. The use of appropriate controls is particularly important for evaluating sediment contamination (see Section 11).

4.2.4 Determine which statistical procedures should be applied to analysis of the data, and define the limits of applicability of the resultant analyses in data interpretation (Test Method E1706).

## 5. Summary of Guide

5.1 This guide provides general guidance and objectives for conducting biological tests with sediments. Detailed technical information on the conduct and evaluation of specific sediment tests is included in other documents referenced in this guide.

5.2 Neither this guide nor any specific test methodology can adequately address the multitude of technical factors that must be considered when designing and conducting a specific investigation. The intended use of this document is therefore not to provide detailed guidance, but rather to assist the investigator in developing technically sound and environmentally relevant biological tests that adequately address the questions being posed by a specific investigation.

## 6. Significance and Use

6.1 Contaminated sediments may affect natural populations of aquatic organisms adversely. Sediment-dwelling organisms may be exposed directly to contaminants by the ingestion of sediments and by the uptake of sediment-associated contaminants from interstitial and overlying water. Contaminated sediments may affect water column species directly by serving as a source of contaminants to overlying waters or a sink for contaminants from overlying waters. Organisms may also be affected when contaminated sediments are suspended in the water column by natural or human activities. Water column species and nonaquatic species may also be affected indirectly by contaminated sediments by the transfer of contaminants through ecosystems (7, 8).

6.2 The procedures described in this guide may be used and adapted for incorporation in basic and applied research to determine the ecological effects of contaminated sediments. These same methods may also be used in the development and implementation of monitoring and regulatory programs designed to prevent and manage sediment contamination.

6.3 Sediment tests with aquatic organisms can be used to quantify the acute and chronic toxicity and the bioavailability of new and presently used materials. Sediment toxicity may also result from environmental processes such as ammonia generation, pH shifts, or dissolved oxygen fluctuation. In many cases, consideration of the adverse effects of sediment-associated contaminants is only one part of a complete hazard assessment of manufactured compounds that are applied directly to the environment (for example, pesticides) and those released (for example, through wastewater effluents) as by-products from the manufacturing process or from municipalities (7).

6.4 Sediment tests can be used to develop exposure-response relationships for individual toxicants by spiking clean sediments with varying concentrations of a test chemical and determining the concentration that elicits the target response in the test organism (Guide E1391). Sediment tests can also be designed to determine the effects that the physical and chemical properties of sediments have on the bioavailability and toxicity of compounds.

6.5 Sediment tests can provide valuable information for making decisions regarding the management of contaminated sediments from hazardous waste sites and other contaminated areas. Biological tests with sediments can also be used to make defensible management decisions on the dredging and disposal of potentially contaminated sediments from rivers and harbors. (7, 8), Test Method E1706.)

## 7. Hazards

### 7.1 General Precautions:

7.1.1 Development and maintenance of an effective health and safety program in the laboratory requires an ongoing commitment by laboratory management and includes: (1) the appointment of a laboratory health and safety officer with the responsibility and authority to develop and maintain a safety program, (2) the preparation of a formal, written health and safety plan, which is provided to each laboratory staff member, (3) an ongoing training program on laboratory safety, and (4) regular safety inspections.

7.1.2 Collection and use of sediments may involve substantial risk to personal safety and health. Chemicals in field-collected sediment may include carcinogenics, mutagens, and other potentially toxic compounds. Inasmuch as sediment testing is often started before chemical analysis can be completed, worker contact with sediment needs to be minimized by (1) using gloves, laboratory coats, safety glasses, face shields and respirators as appropriate, (2) manipulating sediments under a ventilated hood or in an enclosed glove box, and (3) enclosing and ventilating the exposure system. Personal collecting sediment samples and conducting tests should take all safety precautions necessary for the prevention of bodily injury and illness which might result from ingestion or invasion of infectious agents, inhalation or absorption of corrosive or toxic substances through skin contact, and asphyxiation because of lack of oxygen or presence of noxious gases.

7.1.3 Before beginning sample collection and laboratory work, personnel should determine that all the required safety equipment and materials have been obtained and are in good condition.

### 7.2 Safety Equipment:

7.2.1 *Personal Safety Gear*—Personnel should use safety equipment, such as, rubber aprons, laboratory coats, respirators, gloves, safety glasses, face shields, hard hats, and safety shoes. Before beginning sample collection and laboratory work, personnel should be properly trained in the following: (1) when and what personal protective equipment (PPE) is necessary, (2) How to properly wear PPE, (3) limitations to the PPE, and proper care maintenance, useful life, and (4) disposal of PPE (29 CFR 1910.132(f)).

**7.2.2 Laboratory Safety Equipment**—Each laboratory should be provided with safety equipment such as first-aid kits, fire extinguishers, fire blankets, emergency showers, and eye wash stations. Mobile laboratories should be equipped with a telephone to enable personnel to summon help in case of emergency.

### 7.3 General Laboratory and Field Operations:

**7.3.1** Special handling and precautionary guidance in Material Safety Data Sheets (MSDS) should be followed for reagents and other chemicals purchased from supply houses.

**7.3.2** Work with some sediments may require compliance with rules pertaining to the handling of hazardous material. Personnel collecting samples and performing tests should not work alone.

**7.3.3** It is advisable to wash the exposed parts of the body with bacterial soap and water immediately after collecting or manipulating sediment samples.

**7.3.4** Strong acids and volatile organic solvents should be used in a fume hood or under an exhaust canopy over the work area.

**7.3.5** An acidic solution should not be mixed with a hypochlorite solution because hazardous fumes might be produced.

**7.3.6** To prepare and dilute acid solutions, concentrated acid should be added to water, not vice versa. Opening a bottle of concentrated acid and adding concentrated acid to water should be performed only under a fume hood.

**7.3.7** Use of ground-fault systems and leak detectors is strongly recommended to help prevent electrical shocks. Electrical equipment or extension cords not bearing the approval of Underwriter Laboratories should not be used. Ground-Fault interrupters should be installed in all “wet” laboratories where electrical equipment is used.

**7.3.8** All containers should be adequately labeled to indicate their contents.

**7.3.9** A clean well-organized work place contributes to safety and reliable results.

**7.4 Disease Prevention**—Personnel handling samples which are known or suspected to contain human wastes should be immunized against hepatitis B, tetanus, typhoid fever and polio. Thorough washing of exposed skin with bacterial soap should follow handling of samples collected in the field.

**7.5 Safety Manuals**—For further guidance on safe practices when handling sediment samples and conducting toxicity tests, check with the permittee and consult general industrial safety manuals including (9, 10).

**7.6 Pollution Prevention, Waste Management and Sample Disposal**—Guidelines for the handling and disposal of hazardous material should be strictly followed (Guide D4447). The Federal Government has published regulations for the management of hazardous waste and has given the States the option of either adopting those regulations or developing their own. If States develop their own regulations they are required to be as stringent as the Federal regulations. As a handler of hazardous materials, it is your responsibility to know and comply with the

pertinent regulations applicable in the State in which you are operating. Refer to (11) for the citations of the Federal requirements.

## 8. Sediment Test Types

**8.1** Many methods for assessing the toxicity of saltwater and freshwater sediments to benthic organisms have been reported. Those methods are provided in Table 1 for saltwater tests and in Table 2, for freshwater tests, respectively.

**8.2** The selection of a specific toxicity test type is intimately related to the objectives of the sediment evaluation program. These assessments, whether they be for monitoring, regulatory, or research purposes, should be guided by a set of null hypotheses that define the appropriate exposure route and the endpoint of interest.

**8.3** Organism exposure methods most commonly employ the whole sediment in the bedded phase (solid phase), but pore water, suspended and elutriate phase exposures have also been used (7).

**8.4** Programs seeking to characterize or rank sediments on a basin-wide or regional scale typically use whole sediment, solid-phase exposures. Regulatory or permitting programs for dredged material disposal at a containment site may also evaluate this exposure route (8, 12). Disposal at a dispersive site, or concerns over the resuspension and transport of in-place sediments, would suggest the use of suspended phase or elutriate exposures (Annex A1).

**8.5** Methods have been developed to isolate and test the toxicity of elutriates (99) or sediment interstitial water (100) to aquatic organisms. The elutriate test was developed for assessing the potential acute effects of open-water disposal of dredged material. Tests with elutriate samples are used to estimate the water-soluble constituents that may be released from sediment to the water column during disposal operations (101). Toxicity tests of the elutriate with water column organisms have generally indicated that little toxicity is associated with the discharge material (4). However, elutriates have been reportedly more toxic than interstitial water samples (102).

**8.5.1** For many benthic invertebrates, the toxicity and bioaccumulation of sediment-associated contaminants, such as metals and non-ionic organic contaminants, may be correlated with the concentration of these chemicals in the interstitial water (100, 103). The sediment interstitial water toxicity test was developed for assessing the potential *in situ* effects of contaminated sediment on aquatic organisms. Once the interstitial water (or elutriate) has been isolated from the whole sediment, the toxicity testing procedures are similar to effluent toxicity testing with non-benthic species. If benthic species are used as test animals, they may be stressed by the absence of sediment (4).

**8.5.2** The examination of organic extracts may have specific uses. However, caution should be exercised in the use of organic extracts since the availability of sediment contaminants to organisms may have been altered (7).

## 9. Biological Responses

**9.1** Toxicity endpoints in sediment tests range from lethality, growth, reproductive impairment, and physiological

**TABLE 1 Organisms Used in Assessing the Toxicity of Saltwater Sediments<sup>A</sup>**

Taxa	Exposure	Reference
<i>Mortality</i>		
Amphipods	So <sup>B</sup>	(12,13, 14, 15-20), Guide E1367
	Su <sup>C</sup>	(21, 20-24)
Bivalves	So	(16, 20) Guide E724
	Su	(20,25,26)
Copepods	So	(15)
	Su	(15)
Crab	Su	(26)
Cumaceans	So	(14, 17-19)
Fish	El <sup>D</sup>	(27,28)
	So	(20,29)
	Su	(20,25)
Isopods	So	(15)
	Su	(15)
Lobster	Su	(26)
Mysids	So	(20)
	Su	(20-24)
Polychaetes	So	(16,30,31) Guide E1611
Phytoplankton	El	(32)
Shrimp	So	(15, 30-34)
	Su	(15,28,33,34)
Tunicate	Su	(26)
<i>Avoidance/behavior</i>		
Amphipods	So	(35,36)
Bivalves	So	(35,37,38-40)
Crab	So	(35,36)
Echinoderm	So	(35)
Fish	So	(37,41)
Lobster	So	(35)
Polychaetes	So	(37,42)
Shrimp	So	(35,37)
<i>Growth/reproduction/life cycle</i>		
Amphipods	Su	(24)
Bivalves	Su	(43) Guide E724
Copepods	So	(44)
Fish	Su	(45)
Mysids	Su	(22,23,46)
Nematodes	So	(47)
Polychaetes	So	(45,48,49) Guide E1611
	Su	(45,48,49)
Sea urchin	El	(50) Guide E1563 E1525-02(2023)
<i>Pathology</i>		
Amphipods	So	(51)
	Su	(51)
Bivalves	So	(51)
	Su	(51)
Fish	So	(27,52,53)
	Su	(52)
Oyster	So	(52)
	Su	(52)
Polychaetes	So	(51)
	Su	(51)
<i>Physiology</i>		
Fish	Su	(54)
Oligochaetes	El	(55)
Polychaetes	So	(48)
	Su	(48,56)
Shrimp	Su	(56)
<i>Chromosome damage</i>		
Fish	El	(57-59)
Polychaetes	Su	(60)
<i>Bacterial activity</i>		
Bacteria	El	(61,62)
<i>Community recolonization</i>		
Macrobenthos	So	(63-69)

<sup>A</sup> Many of these species have a wide salinity tolerance and therefore may be suitable for testing estuarine sediments.

<sup>B</sup> So—solid-phase sediment exposure.

<sup>C</sup> Su—suspended sediment exposure.

<sup>D</sup> El—elutriate, extract, pore water exposure.

**TABLE 2 Organisms Used in Assessing the Toxicity of Freshwater Sediments<sup>A</sup>**

Taxa	Exposure	Reference
<i>Mortality</i>		
Amphipods	El	(70)
	So	(5,6,8,71,70-73) Test Method E1706
Cladocerans	El	(70)
	So	(5,70,72,74-84) Test Method E1706
	Su	(82)
	El	(70)
	So	(70,72, 74-77)
Insect larvae	El	(70)
	So	(5,8,85, 70-81, 86) Test Method E1706
Isopods	So	(74-77)
Oligochaetes	So	(87-89) Guide E1688
<i>Growth/reproduction</i>		
Amphipods	So	(5,6,71) Test Method E1706
Bacteria	El	(90)
	So	(90)
Cladocerans	El	(90) Test Method E1706
	So	(5,90)
Fish	El	(90)
	So	(90)
Insect larvae	So	(85,86,91,92) Test Method E1706
Nematodes	El	(93)
<i>Physiology</i>		
Oligochaetes	El	(94,95)
<i>Genetic damage</i>		
Fish	El	(2,57,58,94,95)
Nematodes	El	(93)
<i>Bacterial activity</i>		
Bacteria	El	(96,97)
<i>Behavior</i>		
Oligochaetes	So	(98)

<sup>A</sup> Many of these species have a salinity tolerance and therefore may be suitable for testing estuarine sediments.

responses to alterations in community levels of organization (Table 1 and Table 2). Selection of the proper toxic endpoint is predicated largely on the objectives of the evaluation program and the available resources, time, and available methods. Several endpoints are suggested in published methods to measure the potential effects of contaminants in sediment including, survival, growth, behavior, or reproduction; however, survival of test organisms in 10-d exposures is the endpoint most commonly reported (Table 1 and Table 2). These short-term exposures which only measure effects on survival can be used to identify high levels of contamination on sediments, but may not be able to identify moderate levels of contamination in sediments (Test Method E1706, (8)). Sublethal endpoints in sediment tests might also prove to be better estimates of responses if benthic communities to contaminants in the field (85-106).

9.2 The decision to conduct short-term or long-term toxicity tests depends on the goal of the assessment. In some instances, sufficient information may be gained by measuring sublethal endpoints in 10-d tests. In other instances, the 10-d test could be used to screen samples for toxicity before long-term tests are conducted. While the long-term tests are needed to determine direct effects on reproduction, measurement of growth in these toxicity tests may serve as an indirect estimate of

reproductive effects of contaminants associated with sediments (Test Method E1706, (8)).

9.3 Use of sublethal endpoints for assessment of contaminant risk is not unique to toxicity testing with sediments. Numerous regulatory programs require the use of sublethal endpoints in the decision-making process (7) including: (1) Water Quality Criteria (and State Standards), (2) National Pollution Discharge Elimination System (NPDES) effluent monitoring (including chemical-specific limits and sublethal endpoints in toxicity tests); (3) Federal Insecticide, Rodenticide and Fungicide Act (FIFRA) and the Toxic Substances Control Act (TSCA, tiered assessment includes several sublethal endpoints with fish and aquatic invertebrates); (4) Superfund (Comprehensive Environmental Response, Compensation and Liability Act, CERCLA); (5) Organization of Economic Cooperation and Development (OECD, sublethal toxicity testing with fish and invertebrates); (6) European Economic Community (EC, sublethal toxicity testing with fish and invertebrates); and (7) the Paris Commission, (behavioral endpoints).

## 10. Test Organisms

10.1 Once the exposure routes and endpoints of interest have been established, several criteria should be considered when selecting appropriate species (3, 8, 107) and Test Method E1706 for which tests can be conducted that have ecologically relevant endpoints. Ideally, the test species should meet the following criteria:

10.1.1 Have a toxicological (sediment) database demonstrating sensitivity to a range of contaminants or the contaminant of interest, and be taxonomically identified;

10.1.2 Be readily available through field collection or culture;

10.1.3 Be easily maintained in the laboratory;

10.1.4 Be ecologically or economically important;

10.1.5 Have a broad geographical distribution, or be indigenous to the site being evaluated or have a similar niche, be in the same feeding guild, or be similar in behavior to an inhabitant (species);

10.1.6 Be tolerant to a broad range of sediment physico-chemical characteristics (for example, organic carbon and grain size);

10.1.7 Be compatible with selected exposures and endpoints; and

10.1.8 Be tolerant of a range of different water quality characteristics.

10.2 Of these criteria, demonstrated sensitivity to contaminants, ecological relevance, and tolerance to varying sediment physico-chemical characteristics are the most important. The sensitivity of a species to contaminants should be balanced with the concept of discrimination. Species responses may need to provide discrimination between different levels of contamination. Additionally, insensitive species may be preferred for determining bioaccumulation potential. The use of indigenous species that are ecologically important and collected easily is often very straightforward; however, many indigenous species at a contaminated site may be insensitive to contaminants (Guide E1688). Indigenous species might present

a greater concern relative to bioaccumulation potential. With the exception of some saltwater amphipods, few test species have broad sediment toxicity databases. Additionally, many species can be maintained in the laboratory long enough for acclimation to test conditions, but very few are cultured easily. Widespread toxicity testing will require cultured organisms or the use of standard source populations that can be transported without experiencing excessive stress.

10.3 Toxicity is related to the species-specific physiological and biochemical response to a toxicant and the degree of contact between the sediment and the organism. Feeding habits, including the type of food and feeding rate, will influence the exposure of contaminants from sediment (108). Infaunal deposit-feeding species can receive an exposure of sediment contaminants by means of three exposure routes: interstitial water, sediment particles, and overlying water. Benthic invertebrates may selectively consume particles with higher organic carbon and higher contaminant concentrations. Organisms in direct contact with sediment may also accumulate contaminants by direct adsorption to the body wall or exoskeleton, or by absorption through the integument (109). Estimates of bioavailability will thus be more complex for epibenthic animals that inhabit both the sediment and the water column. Some benthic species are exposed primarily by detrital feeding (110). Detrital feeders may not receive most of their body burden directly from interstitial water. For certain higher Kow compounds, uptake by the gut can exceed uptake across the gill (111, 112). However, for many benthic invertebrates, the toxicity and bioaccumulation of sediment-associated contaminants such as metals, kepone, fluoranthene, and organochlorines are highly correlated with the concentration of these chemicals in the interstitial water (100).

10.4 The saltwater test species include a broad spectrum of taxa and feeding types including crustaceans, bivalves, polychaetes, and fish (Table 1). Tests using amphipods have received a great deal of attention because of their overall sensitivity and because they are often absent from contaminated sites (13). This sensitivity has led to the development of routine methods using the burrowing amphipod *Rheopoxynius abronius*. This 10-day acute toxicity test has recently been adapted for use with other amphipod species and has been established (Guide E1367, (14,12)). Since 1977, the U.S. Army Corps of Engineers dredging permit program has routinely required tests with three species: a bivalve, a polychaete, and a fish or shrimp, incorporating both species that burrow into the sediment and those which inhabit the water column. Broad applications of these protocols reveal that these tests are not as sensitive as those with amphipods, and the latter have recently been recommended for permit programs.

10.5 Freshwater sediment tests use a number of different species, including amphipods, midges, mayflies, cladocerans, and oligochaetes (Table 2). Whole sediment tests with the amphipod *Hyalella azteca* generally start with juvenile animals and are typically conducted for 10 to 14-d with measurement of survival or growth (Test Method E1706, (8,71)). Methods for conducting 42-d tests with *H. azteca* have been described in

Test Method **E1706** and **(8)**. Endpoints measured in these long-term tests with *H. azteca* include survival, growth, and reproduction.

10.6 Tests with midge *Chironomus tentans* are generally started with second instar larvae (10 to 14 days old) and continued for 10 to 17 days until the fourth instar; larval survival or growth is the measure of toxicity (Test Method **E1706** **(8, 85)**). Methods for conducting 60-d tests with *C. tentans* have been described in Test Method **E1706** and **(8)**. Exposures start with first instar *C. tentans* and endpoints measured in these long-term tests include survival, growth, emergence, reproduction, and egg hatching. Whole sediment testing procedures with the midge *C. riparius* are started with 1 to 3-day-old larvae and may continue through pupation and adult emergence (**(6)** Test Method **E1706**). Midge exposures started with older larvae may underestimate midge sensitivity to toxicants. For instance, first instar *C. tentans* larvae were 6 to 27 times more sensitive than fourth instar larvae to acute copper exposure **(5, 113)**, and first instar *C. riparius* larvae were 127 times more sensitive than second instar larvae to acute cadmium exposure **(114)**.

10.7 Sediment toxicity tests with mayflies and cladocerans are generally conducted for up to 10 days **(5, 115, 116)** and Test Method **E1706**. Survival and molting frequency are the toxicity endpoints monitored in the mayfly tests, and survival, growth, and reproduction are monitored in the cladoceran tests. While cladocerans are not in direct contact with the sediment, they are frequently in contact with the sediment surface and are probably exposed to both water-soluble and particulate bound contaminants in the overlying water and surface sediment (Test Method **E1706**). Cladocerans are also one of the more sensitive groups of species used in aquatic toxicity testing.

10.8 The most frequently described sediment testing procedures for oligochaetes are acute toxicity testing methods **(98, 8)** also see, Guide **E1688**. However, methods for conducting up to 500-day oligochaete exposures, with growth and reproduction as the toxicity endpoints, have been described **(117)**. A shorter 28-d test starting with sexually mature *Tubifex tubifex* has been described **(118)**. Effects on growth and reproduction are monitored in this shorter test, and the duration of the exposure makes the test more useful for routine sediment toxicity assessments with oligochaetes (Test Method **E1706**). Many oligochaetes have complex life cycles and reproductive strategies, and therefore laboratory culturing requirements have prohibited their use in toxicity testing **(119)**. However, culturing procedures have been described for *Lumbriculus variegatus* and *Tubifex tubifex* **(8, 120,121)** (See also, Test Method **E1706** and Guide **E1688**).

10.9 Because of the database that has been developed with existing tests, it is recommended that, for whole sediment exposures, either phoxocephalid, ampeliscid, or haustoriid amphipods be used in saltwater tests. For freshwater applications, hyalellid amphipods, midge larvae, or mayfly larvae would be appropriate. As new methods are developed, it will be important to establish the sensitivity of each method relative to a benchmark procedure for comparative purposes **(2)**. The whole sediment benchmark for saltwater tests should

be the *Rheopoxynius abronius* survival 10-day acute test, and for freshwater tests it should be *Hyalella azteca* survival and growth in 28-d exposures **(122)**. While chronic tests with whole sediments have been described for a variety of freshwater tests, research is ongoing to describe chronic tests with marine amphipods.

10.10 Multispecies and microcosm tests can also be used to evaluate potential ecosystem responses to contaminated sediments. The use of multi-species tests may provide toxicity information not available from single-species tests since relative species sensitivity may vary among contaminants **(6)**. However, results from multi-species or microcosm tests are more difficult to interpret due to interactions and limited reference literature **(123, 124)**.

## 11. Experimental Design Considerations

### 11.1 Sampling Methods:

11.1.1 Sampling methods are dependent on the purpose and design of the study. The probable source and type of contamination and the objectives of the study should be evaluated before developing a sediment sampling regime. The number and type of samples taken depends on the objectives of the study **(125-128)**.

11.1.2 The number of replicate samples taken at a site should be determined based on the objectives of the study and a preliminary survey of sediment variability at the site. Information from the preliminary survey and the objectives of the study can be used to determine the minimum number of replicates that should be sampled at each site **(126, 127)**.

11.1.3 In general, both toxicity and bioaccumulation tests require at least two exposures: a control and one or more test treatments (see **11.3.12**). The experimental unit for each test is the exposure chamber. A sediment sample is typically split into four or more test chambers. Individual observations obtained from within an individual chamber should not be used as replicate observations. Replicate chambers for a particular sediment provide an estimate of the variability within the test system and are not considered sediment sample or location replicates.

11.1.4 There are several acceptable methods of sampling sediments, for example, corers and grabs or dredges. Grabs or dredges (for example, Ponar or Ekman) are appropriate when sediments are known to be unstratified with respect to the contaminants of concern. If the contaminants are in strata, or if their accumulation rates are of interest, one of several core samplers should be used.  $Pb^{210}$  or  $Cs^{137}$  dating can be performed on cores to identify the thickness of the mixed layer **(125, 128)**. See Guide **E1391** for additional details.

### 11.2 Sample Handling:

11.2.1 Sample handling and preservation are discussed in Guide **E1391** and Test Method **E1706**, and depend on the type of chemical characterization that will be performed. Any sediment disturbance may alter the chemical characterization of that sediment from *in situ* conditions. The use of clean sampling devices and sample containers is essential to ensure the accurate determination of sediment contamination **(126, 128)**.

11.2.2 Physical and chemical characterization of sediments is highly dependent on the needs of the investigator, but it may include loss on ignition, percent water, grain size, total organic carbon, total phosphorus, nitrogen forms, trace metals and organic compounds, pH, total volatile solids, biological oxygen demand, chemical oxygen demand, cation exchange capacity, Eh, pE, total inorganic carbon, acid volatile sulfides, and ammonia (125, 127, 128). Many times, a sediment of concern has some historical data that are used as a basis for selection.

11.2.3 Indigenous organisms may be present in field-collected sediments. An abundance of the same organism or organisms taxonomically similar to the test organism in the sediment sample may make interpretation of treatment effects difficult. Previous investigators have inhibited the biological activity of sediment with sieving, heat, mercuric chloride, antibiotics, or gamma irradiation. (Guide E1391.) However, further research is needed to determine effects on contaminate bioavailability or other modifications of sediments from treatments such as those used to remove or destroy indigenous organisms.

11.2.4 Field-collected sediment samples tend to settle during shipment. As a result, water above the sediment should not be discarded, but should be mixed back into the sediment during homogenization (Test Method E1706). Sediment samples should not be routinely sieved to remove indigenous organisms unless there is a good reason to believe they will influence the response of the test organisms. Large indigenous organisms and large debris can be removed using forceps. Reynoldson et al. (129), observed reduced growth of amphipods, midges, and mayflies in sediments with elevated numbers of oligochaetes and recommended sieving sediments suspected to have high numbers of indigenous oligochaetes. One approach might be to sieve an aliquot of each sediment before the start of a test. If potential predators are recovered from a sediment, it may be desirable to sieve all of that sample before the start of the test. Depending on the objective of the test, it may be necessary to sieve all sediments or run a sieved and un-sieved treatment in parallel to account for potential effects of sieving on test results and subsequent comparisons. The size of the sieve used will depend on the size of the organisms in the sediment sample. If a sediment must be sieved, it is desirable to analyze a sample before and after sieving (for example, measure pore-water metals, dissolved organic carbon (DOC), acid volatile sulfide (AVS), total organic carbon (TOC)) to document the influence of sieving on sediment chemistry.

### 11.3 Exposure Design:

11.3.1 In addition to being available in adequate supply, overlying water used in toxicity tests, and water used to hold organisms before testing, should be acceptable to the test species and uniform in quality. To be acceptable the water must allow the test species to survive and grow without showing signs of disease or apparent stress, such as discoloration or unusual behavior.

11.3.2 Natural overlying water should be uncontaminated and of constant quality and should meet the specifications established in Guide E729. Water should be characterized in accordance with Guide E729 at least twice each year and more

often if (1) such measurements have not been determined semiannually for at least two years or (2) surface water is used.

11.3.3 A natural overlying water is considered to be of uniform quality if the monthly ranges of hardness and alkalinity are less than 5 mg/L or 10 % of their respective averages, whichever is higher, and if the monthly range of pH is less than 0.4 units. Natural overlying waters should be obtained from an uncontaminated well or spring, if possible, or from a surface water source. If surface water is used, the intake should be positioned to minimize fluctuations in quality and the possibility of contamination and maximize the concentration of dissolved oxygen and to help ensure low concentrations of sulfide and iron. For sediment studies with saltwater, the range of salinity should be less than 10 % of the average. In addition, the ion concentrations of the water should be within 10 % of the ion concentrations (adjusted for the salinity) listed in Guide E729. Chlorinated water should not be used for, or in the preparation of, overlying water because residual chlorine and chlorine-produced oxidants are toxic to many aquatic animals and dechlorination is often incomplete.

11.3.4 For certain applications, the experimental design might require the use of water from the test sediment collection site.

11.3.5 Reconstituted fresh and salt water is prepared by adding specified amounts of reagent grade chemicals to high-quality distilled or deionized water (see Guide E729 and Test Method E1706). Acceptable water can be prepared using deionization, distillation, or reverse-osmosis units. Conductivity, pH, hardness, and alkalinity should be measured on each batch of reconstituted water. If the water is prepared from a surface water, the total organic carbon or chemical oxygen demand should be measured on each batch. Filtration through sand, rock, bag, or depth-type cartridge filters may be used to keep the concentration of particulate matter acceptably low. The reconstituted water should be intensively aerated before use, except that buffered soft fresh waters should be aerated before, but not after, the addition of buffers. Problems have been encountered with some species in some fresh reconstituted waters, but these problems can be overcome by aging the reconstituted water for one or more weeks (Guide E729).

11.3.6 Materials used to construct test chambers may include glass, stainless steel, silicone, plastics, and fiberglass that have been prepared properly and tested for toxicity (Guides E1367 and Test Method E1706). The materials selected to construct test chambers may differ, depending on the types of contaminants in the sediments. Within a test, chambers need to be of the same material.

11.3.7 The use of site water or reconstituted water in toxicity tests may depend on the type of test to be performed and the time lapse between sample collection and start of the test.

11.3.8 Static sediment toxicity tests are the simplest to perform and have been used commonly. In such tests, water overlying the sediment is not changed during the test period, but it may be added to replace that which has evaporated. Since



changes in water quality may affect the availability of contaminants to the test species, static exposures are more appropriate for acute tests (7 to 10 days).

11.3.9 Flow-through exposure chambers are suggested for use in chronic tests or with larger animals. Since water is renewed on a continual basis, fewer water quality changes are likely due to the buildup of waste products or interactions between the sediment and overlying water. Flow-through exposures may bias the results of the test by either encouraging the continual release of water-soluble contaminants throughout the test, or by depleting water-soluble contaminants from the sediment early in the test.

11.3.10 General water quality (variables such as pH, salinity, dissolved oxygen, ammonia, and temperature) in the test chambers should meet culture and maintenance requirements for the test species. These parameters should be monitored and recorded on a frequency appropriate to the test length. For example, if the test duration is only a few days, daily monitoring should be performed. However, if the test will continue for weeks or months, measurements may be reduced to every other day or every few days.

11.3.11 The depth of sediment in test chambers may vary depending on the species being tested, its size and degree of burrowing activity, and its sediment processing rate. The latter should be determined prior to the beginning of a sediment toxicity test (126).

11.3.12 Sediment tests includes a control sediment, (sometimes called a negative control). A control sediment is a sediment that is essentially free of contaminants and is used routinely to assess the acceptability of a test and is not necessarily collected near the site of concern. Any contaminants in control sediment are thought to originate from the global spread of pollutants and do not reflect any substantial inputs from local or non-point sources. Comparing test sediments to control sediments is a measure of the toxicity of a test sediment beyond inevitable background contamination and organism health. A control sediment provides a measure of test acceptability, evidence of test organism health, and a basis for interpreting data obtained from the test sediments. A reference sediment is collected near the area of concern and is used to assess sediment conditions exclusive of materials(s) of interest. Testing a reference sediment provides a site-specific basis for evaluating toxicity (Test Method E1706, (8)). (1) In general, the performance of test organisms in the negative control is used to judge the acceptability of a test, and either the negative control or reference sediment may be used to evaluate performance in the experimental treatments, depending on the purpose of the study. Any study in which organisms in the negative control do not meet performance criteria must be considered questionable because it suggests that adverse factors affected the response of test organisms. Key to avoiding this situation is using only control sediments that have demonstrated record of performance using the same test procedure. This includes testing of new collections from sediment sources that have previously provided suitable control sediment. (2) Because of the uncertainties introduced by poor performance in the negative control, such studies should be repeated to insure accurate results. However, the scope or sampling associated

with some studies may make it difficult or impossible to repeat a study. Some researchers have reported cases where performance in the negative control is poor, but performance criteria are met in a reference sediment included in the study design. In these cases, it might be reasonable to infer that other samples that show good performance are probably not toxic; however, any samples showing poor performance should not be judged to have shown toxicity, since it is unknown whether the adverse factors that caused poor control performance might have also caused poor performance in the test treatments. (3) Natural physico-chemical characteristics such as sediment texture may influence the response of test organisms (Guide E1367). The physico-chemical characteristics of test sediment need to be within the tolerance limits of the test organism. Ideally, the limits of a test organism should be determined in advance; however, controls for factors including grain size and organic carbon can be evaluated if the limits are exceeded in a test sediment. If the physico-chemical characteristics of a test sediment exceed the tolerance range of the test organism, a control sediment encompassing these characteristics can be evaluated. The effects of sediment characteristics on the results of sediment tests can be addressed with regression equations. The use of formulated sediment can also be used to evaluate physico-chemical characteristics of sediment on test organisms (Guide E1367, Test Method E1706) (4) The experimental design depends on the purpose of the study. Variables that need to be considered include the number and type of control sediments, the number of treatments and replicates, and water quality characteristics. For instance, the purpose of the study might be to determine a specific endpoint such as an LC50 and may include a control sediment, a positive control, a solvent control, and several concentrations of sediment spiked with chemical (Test Method E1706).

11.3.13 Test temperature should be chosen based on conditions of particular interest or to match the conditions at the sample site. In either case, the choice of temperature and test species should be compatible.

11.3.14 Dissolved oxygen in overlying water should be maintained between 40 and 100 % saturation.

11.3.15 Light quality (including wavelength composition) and daylength are important because of their impacts on both chemical degradation and organism health. Light should be provided from cool-white fluorescent lamps at an intensity appropriate for the test species.

11.3.16 The photoperiod can be selected to mimic that experienced at the sample site, or to simulate a particular season. Suggested periods of daylight and darkness include 16 h light/8 h dark, 14 h light/10 h dark, 12 h light/12 h dark, 24 h light/0 h dark, or 0 h light/24 h dark. Selection should be based on test needs and species.

11.3.17 Whether test organisms should be fed during the test depends on the test duration and type of test species in use. The addition of food can complicate the interpretation of test results because it adds new particulate material, and the food may interact in unknown ways with contaminants in the sediments (126). Additionally, feeding uncontaminated food may reduce exposure. For acute tests ( $\leq 1$  week), most organisms can survive without being fed. If the species process

sediments directly, and enough sediment has been provided to ensure adequate nutrition, feeding may not be necessary. If the species are fish or filter feeders, food may be required, especially during long tests. If organisms are fed during a sediment test, the excess food is typically not removed.

11.3.18 Test water and sediments should be analyzed for contaminants of concern if the objectives of the study are to determine the sources and concentrations of contaminants. If the test is designed to assess toxicity only, the identification of sources of toxicity is not necessary.

11.3.19 Analyses of specific contaminants in tissues of the test species are necessary if bioaccumulation is of interest. If the measurement of organic chemicals, metals, or other contaminants is desirable, appropriate preservation methods should be followed when the samples are collected.

## 12. Data Interpretation

12.1 Data interpretation must be considered in the initial stages of designing an experimental protocol for a specific investigation. Researchers must be aware that all aspects of an experimental protocol, including sampling techniques, number of test replicates, exposure routes, statistical methods, and selection of test species, will place constraints on data interpretation. Data interpretation must be consistent with the goal of the research program and experimental protocol to ensure the ecological significance and environmental relevance of the results of a specific investigation.

12.2 Bioaccumulation and toxicity of sediment-associated contaminants are important to the individuals of a particular species, however, interpreting the ecological significance of those data are difficult to evaluate ((61) see also, Guide E1688 and Test Method E1706). Toxic effects observed in laboratory exposures may not reflect effects on natural populations. However, bioaccumulation of a contaminant, or a toxic response when compared to that same response in a population exposed to a control sediment, is often undesirable.

12.2.1 Swartz et al. (13) evaluated sediment quality conditions along a sediment contaminated gradient of total DDT using information from 10-d toxicity tests with benthic amphipods, sediment chemistry, and the abundance of benthic amphipods in the field. Survival of amphipods, (*Eohaustorius estaurius*, *Rhepoxynius abronius*, and *H.azteca*) in laboratory toxicity tests was positively correlated to the abundance of amphipods in the field and negatively correlated to total DDT concentrations. The toxicity threshold for amphipods in 10-d sediment toxicity test was about 300 ug total DDT/g organic carbon. The threshold for reduction in abundance of amphipod in the field was about 100 ug total DDT/g organic carbon. Therefore, correlations between toxicity contamination, and the status of benthic macroinvertebrates in the field indicate that 10-d sediment toxicity tests can provide a reliable indicator of the presence of adverse levels of sediment contamination in the field. However, these short-term toxicity tests may be under protective of sublethal effects of contaminants in benthic communities in the field.

12.2.2 Similarly, Canfield et al. (104, 105, 106) evaluated the composition of benthic invertebrate communities in sediments in a variety of locations including the Great Lakes, the

upper Mississippi River, and the Clark Fork River in Montana. Results of these benthic invertebrate community assessments were compared to sediment quality guidelines (SQGs) and 28-d sediment toxicity tests with *H. azteca*. Good concordance was evident between measures of laboratory toxicity, SQGs, and benthic invertebrate composition in extremely contaminated samples. However, in moderately contaminated samples, less concordance was observed between the composition of the benthic community and either laboratory toxicity test or SQGs. The laboratory toxicity tests better identified chemical contamination in sediments compared to many of the commonly used measures of benthic invertebrate community structure. As the status of benthic invertebrates communities may reflect other factors such as habitat alteration in addition to effects of contaminants, the use of longer-term toxicity tests in combination with SQGs may provide a more sensitive and protective measure of potential toxic effects of sediment contamination on benthic communities compared to use of 10-d toxicity tests.

12.2.3 Numerical SQGs have been developed by a variety of federal, state, and provincial agencies across North America using matching sediment chemistry and biological effects data. These SQGs have been routinely used to interpret historical data, identify potential problem chemicals or areas at a site, design monitoring programs, classify hot spots and rank sites, and make decisions for more detailed studies (130, 131, 132, 103) Additional suggested uses for SQGs include identifying the need for source controls of problem chemicals before release, linking chemical sources to sediment contamination, triggering regulatory action, and establishing target remediation objectives (8). Numerical SQGs, when used with other tools such as sediment toxicity tests, bioaccumulation, and benthic community surveys, can provide a powerful weight of evidence for assessing the hazards associated with contaminated sediments (7).

12.3 The calculation procedure(s) and interpretation of the results should be appropriate to the experimental design. Statistical procedures used to calculate test results can be divided into two categories: those that test hypotheses and those that provide point estimates. No procedure should be used without careful consideration of (1) the advantages and disadvantages of various alternative procedures and (2) appropriate preliminary tests, such as those for outliers and heterogeneity (Test Method E1706).

12.4 When samples from field sites are replicated (that is, separate samples from different grabs taken at the same site), site effects (bioaccumulation and toxicity endpoints) can be compared statistically by a one-tailed t-test, analysis of variance (ANOVA), or regression analysis. Analysis of variance is used to determine whether any of the sites are different from the control. This is a test of the null hypothesis, that no differences exist in effects observed among the sites and controls. If the F-test is not statistically significant ( $P > 0.05$ ), it can be concluded that the effects observed in the sites were not large enough to be detected as statistically significant by the experimental design and hypothesis test used. Non-rejection does not mean that the null hypothesis is true. The amount of effect that occurred should be considered.

12.4.1 All exposure concentration effects (or field sites) can be compared with the control effects by using mean separation techniques such as those explained by Chew orthogonal contrasts, Fisher's methods, Dunnett's procedure, or Williams' method (133, 21). The lowest concentration for which the difference in observed effect exceeds the statistical significant difference is defined as the LOEC (lowest observed effect concentration) for that endpoint. The highest concentration for which the difference in effect is not greater than the statistical significant difference is defined as the NOEC (no observed effect concentration) for that endpoint (133).

12.5 In cases in which serial dilution sediment toxicity studies are conducted, the LC50 (median lethal concentration) or EC50 (median effect concentration) and its 95 % confidence limits should be calculated (when appropriate) on the basis of the following: (1) the measured initial sediment concentrations of test material, if available, or the nominal initial sediment concentrations for static tests; and (2) the average measured sediment concentrations of test material, if available, or the nominal average sediment concentrations for flow-through tests. If other LCs or ECs are calculated, their 95 % confidence limits should also be calculated (see Guide E729).

12.6 Most toxicity tests produce quantal data, that is, counts of the number of responses in two mutually exclusive categories, such as alive or dead. A variety of methods (134) can be used to calculate an LC50 or EC50 and 95 % confidence limits from a set of quantal data that is binomially distributed and contains two or more concentrations at which the percent dead or affected is between 0 and 100. The most widely used are the probit, moving average, Spearman-Kärber, and Litchfield-Wilcoxon methods. The method used should appropriately take into account the number of test organisms per chamber. The binomial test can also be used to obtain statistically sound information on the LC50 or EC50 even when there are less than two effective concentrations between 0 and 100 %, assuming mortalities of 0 and 100 % mortality are observed at two different concentrations. The binomial test provides a range within which the LC50 or EC50 should lie.

13. Keywords

13.1 bioaccumulation; contamination; experimental design; freshwater; saltwater; sediment; toxicity

ANNEX

(Mandatory Information)

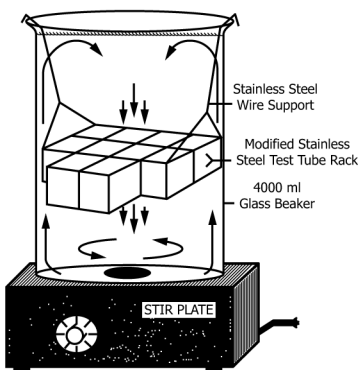
A1. SEDIMENT RESUSPENSION TESTS

A1.1 Scope

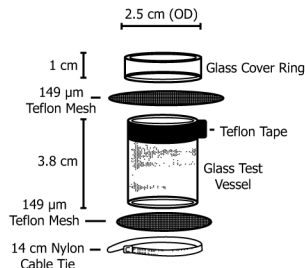
A1.1.1 This annex briefly describes twelve systems for evaluating the effects of suspended solids and their associated contaminants (soluble and insoluble) on aquatic organisms using static, recirculating, or flow-through exposure systems. The main objective, organisms, and apparatus used in these tests are detailed. A brief description of how the apparatus works and any discussion or conclusions reported (see Tables

A1.1-A1.3) for these studies is also included. The following information will strictly provide a general guide to aid future research endeavors.

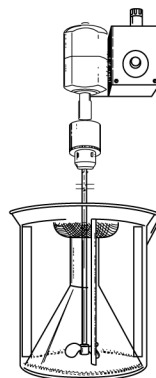
A1.1.2 Sediment suspension and resuspension tests provide information about the bioavailability of contaminants associated with sediments to aquatic organisms. Water column organisms can be exposed to contaminated bottom sediments



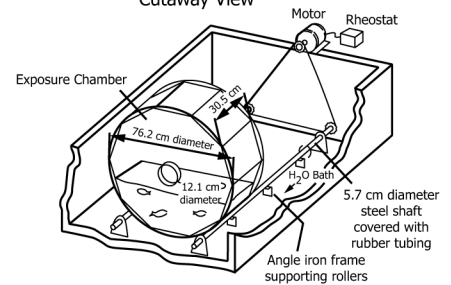
(see Ref. A/)



(see Ref. B/Table A1.1)



WATER BATH  
Cutaway View



(see Ref. D/Table A1.1)<sup>4</sup>

FIG. A1.1 Static/Renewal Tests