SLOVENSKI STANDARD
oSIST prEN ISO 13160:2020

01-april-2020

Kakovost vode - Stroncij Sr-90 in stroncij Sr-89 - Preskusne metode s štetjem s tekočinskim scintilatorjem ali proporcionalnim štetjem (ISO/DIS 13160:2020)

Water quality - Strontium 90 and strontium 89 - Test methods using liquid scintillation counting or proportional counting (ISO/DIS 13160:2020)

Wasserbeschaffenheit - Strontium 90 und Strontium 89 - Verfahren mittels Flüssigszintillationszählung oder Proportionalzählung (ISO/DIS 13160:2020)

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Qualité de l'eau - Strontium 90 et strontium 89 - Méthodes d'essai par comptage des scintillations en milieu liquide ou par comptage proportionnel (ISO/DIS 13160:2020)

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Foreword

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The procedures used to develop this document and those intended for its further maintenance are described in the ISO/IEC Directives, Part 1. In particular, the different approval criteria needed for the different types of ISO documents should be noted. This document was drafted in accordance with the editorial rules of the ISO/IEC Directives, Part 2 (see www.iso.org/directives).

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For an explanation of the voluntary nature of standards, the meaning of ISO specific terms and expressions related to conformity assessment, as well as information about ISO's adherence to the World Trade Organization (WTO) principles in the Technical Barriers to Trade (TBT) see www.iso.org/iso/foreword.html.

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This second edition cancels and replaces the first edition (ISO 13160:2012), which has been technically revised. The main changes compared to the previous edition are as follows:

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Introduction

Radioactivity from several naturally-occurring and anthropogenic sources is present throughout the environment. Thus, water bodies (e.g. surface waters, ground waters, sea waters) can contain radionuclides of natural, human-made, or both origins.

— Natural radionuclides, including ^{40}K, ^{3}H, ^{14}C, and those originating from the thorium and uranium decay series, in particular ^{226}Ra, ^{228}Ra, ^{234}U, ^{238}U, and ^{210}Pb, can be found in water for natural reasons (e.g. desorption from the soil and washoff by rain water) or can be released from technological processes involving naturally occurring radioactive materials (e.g. the mining and processing of mineral sands or phosphate fertilizer production and use).

— Human-made radionuclides, such as transuranium elements (americium, plutonium, neptunium, curium), ^{3}H, ^{14}C, ^{90}Sr, and gamma emitting radionuclides can also be found in natural waters. Small quantities of these radionuclides are discharged from nuclear fuel cycle facilities into the environment as the result of authorized routine releases. Some of these radionuclides used for medical and industrial applications are also released into the environment after use. Anthropogenic radionuclides are also found in waters as a result of past fallout contaminations resulting from the explosion in the atmosphere of nuclear devices and accidents such as those that occurred in Chernobyl and Fukushima.

Radionuclide activity concentration in water bodies can vary according to local geological characteristics and climatic conditions and can be locally and temporally enhanced by releases from nuclear installation during planned, existing and emergency exposure situations[1]. Drinking water may thus contain radionuclides at activity concentrations which could present a risk to human health.

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The radionuclides present in liquid effluents are usually controlled before being discharged into the environment[2]. Water bodies and drinking waters are monitored for their radioactivity content as recommended by the World Health Organization (WHO)[3] so that proper actions can be taken to ensure that there is no adverse health effect to the public. Following these international recommendations, national regulations usually specify radionuclide authorized concentration limits for liquid effluent discharged to the environment and radionuclide guidance levels for water bodies and drinking waters for planned, existing and emergency exposure situations. Compliance with these limits can be assessed using measurement results with their associated uncertainties as specified by ISO/IEC Guide 98-3 and ISO 5667-20[4].

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Depending on the exposure situation, there are different limits and guidance levels that would result in an action to reduce health risk. As an example, during a planned or existing situation, the WHO guidelines for guidance level in drinking water is 100 Bq·l^{-1} for ^{89}Sr activity concentration and 10 Bq·l^{-1} for ^{90}Sr activity concentration.

NOTE 1 The guidance level is the activity concentration with an intake of 2 l/d of drinking water for one year that results in an effective dose of 0,1 mSv/a for members of the public. This is an effective dose that represents a very low level of risk and which is not expected to give rise to any detectable adverse health effects[3].

In the event of a nuclear emergency, the WHO Codex guideline levels[5] mentioned that the activity concentration might not be greater than 1000 Bq·l^{-1} for ^{89}Sr or 100 Bq·l^{-1} for ^{90}Sr for infant food and 1000 Bq·l^{-1} for ^{89}Sr or 100 Bq·l^{-1} for ^{90}Sr for food other than infant food.

NOTE 2 The Codex guidelines levels (GLs) apply to radionuclides contained in food destined for human consumption and traded internationally, which have been contaminated following a nuclear or radiological emergency. These GLs apply to food after reconstitution or as prepared for consumption, i.e. not to dried or concentrated food, and are based on an intervention exemption level of 1 mSv in a year for members of the public (infant and adult)[5].

Thus, the test method can be adapted so that the characteristic limits, decision threshold, detection limit and uncertainties ensure that the radionuclide activity concentrations test results can be verified to be below the guidance levels required by a national authority for either planned/existing situations or for an emergency situation[6][7].

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Usually, the test methods can be adjusted to measure the activity concentration of the radionuclide(s) in either wastewaters before storage or in liquid effluents before discharge to the environment. The test results will enable the plant/installation operator to verify that, before their discharge, wastewaters/ liquid effluent radioactive activity concentrations do not exceed authorized limits.

The test method described in this document may be used during planned, existing and emergency exposure situations as well as for wastewaters and liquid effluents with specific modifications that could increase the overall uncertainty, detection limit, and threshold.

The test method may be used for water samples after proper sampling, sample handling, and test sample preparation (see the relevant part of the ISO 5667 series).

This document has been developed to answer the need of test laboratories carrying out these measurements, that are sometimes required by national authorities, as they may have to obtain a specific accreditation for radionuclide measurement in drinking water samples.

This document is one of a set of International Standards on test methods dealing with the measurement of the activity concentration of radionuclides in water samples.

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Water quality — Strontium 90 and strontium 89 — Test methods using liquid scintillation counting or proportional counting

WARNING — Persons using this document should be familiar with normal laboratory practice. This document does not purport to address all of the safety problems, if any, associated with its use. It is the responsibility of the user to establish appropriate safety and health practices and to determine the applicability of any other restrictions.

IMPORTANT — It is absolutely essential that tests conducted according to this document be carried out by suitably trained staff.

1 Scope

This document provides a selection of test methods and their associated principles for the measurement of the activity of ^{90}Sr in equilibrium with ^{90}Y, and ^{89}Sr, pure beta-emitting radionuclides, in water samples. Different chemical separation methods are presented to produce strontium and yttrium sources, the activity of which are determined using a proportional counter (PC) or liquid scintillation counter (LSC).

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The selection of a particular test method depends on the origin of the contamination, the characteristics of the water to be analyzed, the required accuracy of test results and the available resources of the laboratory.

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$c^{*}_{A,i}$	decision threshold of radionuclide i	Bq l^{-1}
$c^{\#}_{A,i}$	detection limit of radionuclide i	Bq l^{-1}
$c^{\triangleleft}_{A,i}, c^{\triangleright}_{A,i}$	lower and upper limits of the probabilistically symmetric coverage interval of radionuclide i	Bq l^{-1}
$c^{<}_{A,i}, c^{>}_{A,i}$	lower and upper limits of the shortest coverage interval of radionuclide i	Bq l^{-1}
$R_{c,i}$	chemical yield of the extraction of radionuclide i	1
r_0	background count rate	s^{-1}
r_{0j}	background count rate for measurement *j*	s^{-1}
r_g	gross count rate	s^{-1}
r_{gj}	gross count rate for measurement *j*	s^{-1}
r_j	net count rate for measurement *j*	s^{-1}
r_s	calibration source count rate	s^{-1}
t	time elapsed between separation of ^{90}Sr/^{90}Y (t = 0) and mid-point of counting	s
t_0	background counting time	s
t_d, t_f	start and finish time respectively of the measurement, referred to t = 0	s
t_g	sample counting time	s
t_j	start time of the measurement *j*, referred to t = 0	s
t_s	calibration source counting time	s
U	expanded uncertainty, calculated by $U = ku(c_A)$ with k = 1, 2 ...	Bq l^{-1}
$u(c_A)$	standard uncertainty associated with the measurement result	Bq l^{-1}
V	volume of the test sample	l
ε_i	counting efficiency for radionuclide i	1
λ_i	decay constant of radionuclide i	1

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4 Principle

4.1 General

Strontium-90, ^{90}Y and ^{89}Sr are all pure beta-particle emitting radionuclides. Their beta-emission energies and half-lives are given in Table 1.

Table 1 — Half-lives, maximum energies, and average energies of ^{90}Sr, ^{90}Y, and ^{89}Sr [8]

Parameter	^{90}Sr	^{90}Y	^{89}Sr
Maximum energy	546,0 keV	2 283,9 keV	1 491,0 keV
Average energy	196,4 keV	935,3 keV	586,3 keV
Half-life	28,80 (7) a	2,6684 (13) d	50,57 (3) d

Strontium-90 can be directly measured or estimated through the measurement of its decay progeny ^{90}Y. All the test methods are based on a chemical separation step followed by beta-counting of either ^{90}Sr or ^{90}Y using PC or LSC. See Table 2.

4.2 Chemical separation

Strontium is isolated from the water using precipitation, ion exchange or specific chromatographic separation by crown ether resin (see Reference [9]). Yttrium can then be isolated by precipitation or liquid–liquid extraction.

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The method chosen shall be selective with a high chemical yield. When thorium, lead or bismuth radioisotopes are present at high activity levels, they may interfere with ^{90}Sr, ^{90}Y or ^{89}Sr detection. Other matrix constituents such as alkaline earth metals and in particular calcium for strontium, or transuranic and lanthanide elements for yttrium, reduce the chemical yield of the extraction.

The radiochemical separation yield is calculated using a carrier such as stable strontium or yttrium, or a radioactive tracer such as ^{85}Sr. Techniques like atomic absorption spectroscopy (AAS), inductively coupled plasma–atomic emission spectroscopy (ICP–AES) or inductively coupled plasma–mass spectrometry (ICP–MS) to measure the carrier, and gamma-spectrometry to measure ^{85}Sr, are recommended. A carrier can also be measured by gravimetric methods, but the presence of inactive elements, essentially alkaline earth elements, in the leaching solutions can lead to an overestimation of the radiochemical separation yields, particularly for the measurement of strontium.

When stable strontium is added as a carrier, the original strontium concentration in the test sample shall be known to determine the chemical yield.

4.3 Detection

The use of LSC, which provides spectra and permits the detection of interference from unwanted radionuclides, is recommended in preference to PC, which does not distinguish between emissions from different beta-emitters. When PC is used, it is recommended that the purity of the precipitate is checked by following the change over an appropriate time of the ^{90}Y or ^{89}Sr activity, even though this is time consuming.

Six test methods are presented in Annexes A, B, C, D, E, and F.

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5 Chemical reagents and equipment

The necessary chemical reagents and equipment for each strontium measurement method are specified in Annexes A, B, C, D, E, and F.

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During the analyses, unless otherwise stated, use only reagents of recognized analytical grade and laboratory water such as distilled or demineralized water or water of equivalent purity as specified in ISO 3696[10].

6 Procedure

6.1 Test sample preparation

Strontium is determined from the water test sample after appropriate sampling procedures[18][19][20].

If filtration is required, it should be done prior to the addition of the tracer or carrier and sufficient time should be allowed to attain chemical equilibrium before starting the test sample preparation.

When stable strontium is added as a carrier, the original strontium concentration in the test sample shall be known to determine the chemical yield. In the case of separation based on extraction chromatography, the total strontium content shall be below the sorption capacity of the resin to avoid saturation of the resin.

6.2 Chemical separation

6.2.1 General

There are several routine analyses of ^{89}Sr and ^{90}Sr involved in the separation and purification of strontium: precipitation, liquid–liquid extraction or chromatographic techniques (ion exchange or chromatographic extraction). Annexes A, B, C, D, E, and F describe a test method for each of these techniques.

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Table 2 — Determination procedures for strontium depending on its origin

Origin		Old contamination				Fresh contamination	
Radionuclide		$^{90}Sr+^{90}Y$				$^{90}Sr+^{90}Y$ ^{89}Sr	
Separation	Element	Sr		Y[a]		Sr	
	Method	Chromatography[b]	Precipitation	Extraction	Precipitation	Chromatography[b]	Precipitation
	Product	^{90}Sr		^{90}Y		$^{90}Sr+^{90}Y$	
	Carrier or Tracer[c]	^{85}Sr or stable Sr		Stable Y		^{85}Sr or stable Sr	
Measurement(s)	Equilibrium $^{90}Sr+^{90}Y$ 20 d	Yes (recommended)	No	No		Yes	No
	Number	One		One		Two or more	
	Emissions	^{90}Sr ^{90}Y		^{90}Y		^{90}Sr ^{90}Y ^{89}Sr	
	Equipment	PC or LSC (total)		PC or LSC (total or Cerenkov)		PC or LSC (total)	
	Calibration sources	$^{90}Sr+^{90}Y$	^{90}Sr ^{90}Y	^{90}Y		$^{90}Sr+^{90}Y$ ^{89}Sr	^{90}Sr ^{90}Y ^{89}Sr

a Y separation is performed after the ^{90}Sr-^{90}Y equilibrium is established in the test sample.

b Liquid chromatography or specific chromatography using crown ether resin.

c Carrier or tracer element measurements can be taken using gamma-spectrometry for ^{85}Sr (tracer), by gravimetry, atomic absorption spectrometry (AAS), inductively coupled plasma mass spectrometry (ICPMS) for Sr and Y (tracer and/or carrier).

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6.2.2 Precipitation techniques

The precipitation technique is suitable for the separation of most mineral elements, including strontium, in water samples with high mineral salt contents. This technique is very efficient, but not selective for strontium. The use of large quantities of nitric acid and the need to wait for the ^{90}Y to reach equilibrium limit its use.

The addition of fuming nitric acid leads to a strontium precipitate with other interfering elements. Successive dissolution–precipitation cycles concentrate strontium in the precipitate, while yttrium and other elements remain in the supernatant fraction. The most usual procedures lead to a $SrCO_3$ precipitate.

For the test method with ^{90}Sr and ^{90}Y in equilibrium, either the total concentration of yttrium and strontium is directly measured in the precipitate or the yttrium activity is measured after a separation from the strontium. In this latter case, the chemical yield is estimated by the addition of an yttrium carrier to the source before the yttrium separation. The final product is an yttrium precipitate, usually in the form of an oxalate.

In the absence of ^{89}Sr, ^{90}Sr is measured by counting the beta-emission of ^{90}Y or of ^{90}Y and ^{90}Sr in equilibrium.

When ^{89}Sr in the water test sample cannot be neglected, the direct measurement method of strontium at two different times shall be chosen.

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Two precipitation methods are described: Annex A employs PC for ^{89}Sr and ^{90}Sr; Annex B employs LSC for ^{89}Sr and ^{90}Sr.

6.2.3 Liquid–liquid extraction technique

This technique is based on the extraction of ^{90}Y in equilibrium with its radioactive parent ^{90}Sr with an organic solvent. The chemical separation is fast and requires few technical resources. A provisional result may be achieved after 3 d (approximately one yttrium decay period). However, total selectivity of the extraction is not always possible. In the presence of high levels of natural radioactivity, interference may occur, making it difficult to determine very low levels of strontium activity.

Yttrium-90 is extracted from the water test sample fraction using an organic solvent, and then after re-extraction, recovered in the form of an yttrium precipitate. Test methods are presented in Annexes C and F.

After the source preparation, the ^{90}Y is measured by PC (Annex F) or by measuring the Cerenkov radiation from the ^{90}Y with LSC (Annex C).

The absence of other interfering beta-emitters is verified during the decay of ^{90}Y by measuring the decrease in count rate of the ^{90}Y and once the decay is complete, comparing it with the background level activity.

6.2.4 Chromatographic technique

6.2.4.1 Ion exchange resin

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This technique is based on Sr(II) exchange on a cationic resin and is used for separation and purification of strontium in large volume samples. A method is presented in Annex D in which the measurement is carried out with a PC.

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6.2.4.2 Crown ether resin

This technique is based on the selective chromatographic separation of strontium using a specific crown ether resin. A method is presented in Annex E in which the measurement is carried out by LSC.

6.3 Preparation of the source for test

6.3.1 Source preparation for liquid scintillation counter

The strontium or yttrium precipitate is dissolved and mixed with a liquid scintillation cocktail. When the strontium or yttrium is already in solution, it is mixed directly with the liquid scintillator. The volume of the analyzed aliquot depends on the equipment (vial size) and the specific scintillation cocktail used.

The calibration source shall be prepared from a known activity of tracer (^{90}Sr, ^{89}Sr, ^{90}Sr + ^{90}Y or ^{90}Y) with the same geometry and chemical composition as the source to be measured.

The blank source should be prepared following the method chosen starting with a clean test sample (or water).

6.3.2 Source preparation for proportional counter

A strontium or yttrium precipitate is deposited on a filter by filtration or on a stainless steel planchet by direct evaporation.

The filter or planchet size diameter should be similar to the detector size (see Annex A or D).

A calibration source shall be prepared from a known amount of tracer (^{90}Sr, ^{89}Sr, ^{90}Sr + ^{90}Y or ^{90}Y) with the same geometry and chemical composition as the source to be measured.

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