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Standard Practice for Passive Soil Gas Sampling in the Vadose Zone for Source Identification, Spatial Variability Assessment, Monitoring, and Vapor Intrusion Evaluations¹

This standard is issued under the fixed designation D7758; the number immediately following the designation indicates the year of original adoption or, in the case of revision, the year of last revision. A number in parentheses indicates the year of last reapproval. A superscript epsilon (ϵ) indicates an editorial change since the last revision or reapproval.

1. Scope

1.1 *Purpose*—This practice covers standardized techniques for passively collecting soil gas samples from the vadose zone and is to be used in conjunction with Guide D5314.

1.2 *Objectives*—Objectives guiding the development of this practice are: (1) to synthesize and put in writing good commercial and customary practice for conducting passive soil gas sampling, (2) to ensure that the process for collecting and analyzing passive soil gas samples is practical and reasonable, and (3) to provide standard guidance for passive soil gas sampling performed in support of source identification, spatial variability/extent determinations, site assessment, site monitoring, and vapor intrusion investigations.

1.3 This practice does not address requirements of any federal, state, or local regulations or guidance or both with respect to soil gas sampling. Users are cautioned that federal, state, and local guidance may impose specific requirements that differ from those of this practice.

1.4 *Units*—The values stated in SI units are to be regarded as standard. No other units of measurement are included in this standard.

1.5 *This standard does not purport to address all of the safety concerns, if any, associated with its use. It is the responsibility of the user of this standard to establish appropriate safety and health practices and determine the applicability of regulatory limitations prior to use.*

1.6 *This practice offers a set of instructions for performing one or more specific operations. This document cannot replace education or experience and should be used in conjunction with professional judgment. Not all aspects of this practice may be applicable in all circumstances. This ASTM standard is not intended to represent or replace the standard of care by which the adequacy of a given professional service must be judged, nor should this document be applied without consideration of*

a project's many unique aspects. The word "Standard" in the title means only that the document has been approved through the ASTM consensus process.

2. Referenced Documents

2.1 ASTM Standards:²

- D653 Terminology Relating to Soil, Rock, and Contained Fluids
- D1356 Terminology Relating to Sampling and Analysis of Atmospheres
- D2216 Test Methods for Laboratory Determination of Water (Moisture) Content of Soil and Rock by Mass
- D2487 Practice for Classification of Soils for Engineering Purposes (Unified Soil Classification System)
- D3740 Practice for Minimum Requirements for Agencies Engaged in Testing and/or Inspection of Soil and Rock as Used in Engineering Design and Construction
- D4597 Practice for Sampling Workplace Atmospheres to Collect Gases or Vapors with Solid Sorbent Diffusive Samplers
- D5088 Practice for Decontamination of Field Equipment Used at Waste Sites
- D5314 Guide for Soil Gas Monitoring in the Vadose Zone
- D5792 Practice for Generation of Environmental Data Related to Waste Management Activities: Development of Data Quality Objectives
- D6196 Practice for Selection of Sorbents, Sampling, and Thermal Desorption Analysis Procedures for Volatile Organic Compounds in Air
- D6311 Guide for Generation of Environmental Data Related to Waste Management Activities: Selection and Optimization of Sampling Design
- E2600 Guide for Vapor Encroachment Screening on Property Involved in Real Estate Transactions

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² For referenced ASTM standards, visit the ASTM website, www.astm.org, or contact ASTM Customer Service at service@astm.org. For *Annual Book of ASTM Standards* volume information, refer to the standard's Document Summary page on the ASTM website.

2.2 U.S. EPA Methods³

Method 8260C Volatile Organic Compounds by Gas Chromatography/Mass Spectrometry (GC/MS)

Method 8270C Semivolatile Organic Compounds by Gas Chromatography/Mass Spectrometry

Method TO-17 Determination of Volatile Organic Compounds in Ambient Air Using Active Sampling Onto Sorbent Tubes

3. Terminology

3.1 Definitions of Terms Specific to This Standard:

3.1.1 This section provides definitions and descriptions of terms used in or related to this practice. A list of acronyms and a list of symbols are also included. The terms are an integral part of this practice and are critical to an understanding of the practice and its use. Also see Terminology **D653** and **D1356**.

3.1.2 *absorption*, *n*—the penetration of one substance into the inner structure of another.

3.1.3 *active sampling*, *v*—means of collecting a gas-phase substance that uses a mechanical device such as a pump or vacuum-assisted critical orifice to draw air into or through a sampling device.

3.1.4 *adsorption*, *n*—adherence of the atoms, ions, or molecules of a gas or liquid to the surface of another substance (chemisorption).

3.1.5 *ambient air*, *n*—any unconfined portion of the atmosphere; open air.

3.1.6 *background level*, *n*—concentration of a substance that is typically found in ambient air (for example, as a result of industrial or automobile emissions), indoor air (for example, from building materials or indoor activities), or the natural geology of an area.

3.1.7 *blank sample*, *n*—clean sample or a sample of matrix processed to measure artifacts in the measurement process.

3.1.7.1 *Discussion*—Blank samples are named according to their type and use (for example, method blank, trip blank, field blank, and preparation or manufacturing blank).

3.1.8 *contaminant*, *n*—substances not normally found in an environment at the observed concentration.

3.1.9 *desorption*, *n*—the process of freeing from a sorbed state.

3.1.10 *duplicate samples*, *n*—two samples taken from and representative of the same population that are carried through all steps of the sampling and analytical procedures in an identical manner.

3.1.11 *field blank*, *n*—clean sampling media that is carried to the sampling site, exposed to ambient air during field sampling procedures, and transported to the laboratory for analysis (also referred to as an ambient air control sample).

3.1.12 *groundwater*, *n*—part of the subsurface water that is in the saturated zone.

3.1.13 *method blank*, *n*—quality control check to measure laboratory contamination during sample analysis.

3.1.14 *moisture content*, *n*—the moisture present in a material, as determined by definite prescribed methods, expressed as a percentage of the mass of the sample on either of the following bases: (1) original mass ; (2) moisture-free (oven dried) mass (see Test Method **D2216**).

3.1.15 *passive sampling*, *v*—means of collecting a gas-phase substance that uses sorbent materials in a sampling device exposed for a finite duration in the medium being sampled.

3.1.16 *preparation blank*, *n*—quality control check to define the efficiency of conditioning a batch of sorbent samplers at the laboratory for sample collection (also referred to as manufacturing blanks).

3.1.17 *porosity*, *n*—volume fraction of rock or soil not occupied by solid material but usually occupied by liquids, vapor, or air, or combinations thereof.

3.1.17.1 *Discussion*—Porosity is the void volume of soil or rock or both divided by the total volume of soil or rock or both.

3.1.18 *sampling rate*, *n*—the ratio of mass of a given compound collected by a diffusive sampler per unit time of exposure to the concentration of that compound in the atmosphere being sampled. The sampling rate is sometimes referred to as the uptake rate.

3.1.19 *saturated zone*, *n*—zone in which all of the voids in the rock or soil are filled with water at a pressure that is greater than atmospheric.

3.1.19.1 *Discussion*—The water table is the top of the saturated zone in an unconfined aquifer.

3.1.20 *soil gas*, *n*—vadose zone atmosphere; soil gas is the air existing in void spaces in the soil between the groundwater table and the ground surface.

3.1.21 *soil moisture*, *n*—water contained in the pore spaces in the vadose zone.

3.1.22 *sorbent*, *n*—a solid or liquid medium in or upon which materials are collected by adsorption, absorption, or chemisorption.

3.1.23 *sorbent sampling*, *v*—the collection of chemicals from an air or emission sample by allowing the air or emissions to contact a sorbent.

3.1.24 *sorption*, *n*—a process by which one material (the sorbent) takes up and retains another material (the sorbate) by the processes of adsorption or absorption.

3.1.25 *source*, *n*—area(s) at a site where releases have occurred that are emanating vapors from either the vadose zone or groundwater.

3.1.25.1 *Discussion*—There may be multiple sources at a site and the area over which any one source is defined is subject to interpretation from multiple data sets.

3.1.26 *spatial variability*, *n*—relationship of organic compound mass from one location to many others at a site as a function of distance.

3.1.27 *starvation effect*, *n*—when the analyte uptake rate of a passive sorbent sampler is greater than the replenishment rate of the analyte around the sampler, which results in a low bias measurement.

³ Available from United States Environmental Protection Agency (EPA), Ariel Rios Bldg., 1200 Pennsylvania Ave., NW, Washington, DC 20004, <http://www.epa.gov>.

3.1.28 *trip blank, n*—clean, unused sampling media that is carried to the sampling site and transported to the laboratory for analysis without having been exposed to field sampling procedures.

3.1.29 *vadose zone, n*—hydrogeological region extending from the soil surface to the top of the principal water table.

3.1.29.1 *Discussion*—Perched groundwater may exist within this zone.

3.1.30 *vapor intrusion, n*—migration of a volatile chemical(s) from subsurface soil or water into an overlying or nearby building.

3.1.31 *water table, n*—top of the saturated zone in an unconfined aquifer.

3.2 Acronyms:

3.2.1 *BLS*—Below land surface (also known as below ground surface (bgs))

3.2.2 *COC*—Compound of concern

3.2.3 *EPA*—Environmental Protection Agency

3.2.4 *ID*—Identification

3.2.5 *MDL*—Method detection limit

3.2.6 *QA/QC*—Quality assurance and quality control

3.2.7 *PSG*—Passive soil gas

3.2.8 *SVOC*—Semivolatile organic compound

3.2.9 *TD-GC/MS*—Thermal desorption-gas chromatography/mass spectrometry

3.2.10 *U.S.*—United States

3.2.11 *VOC*—Volatile organic compound

3.3 Symbols:

3.3.1 *cm*—centimeter

3.3.2 *m*—meter

3.3.3 *mm*—millimeter

3.3.4 *min*—minute

3.3.5 *ng*—mass in nanograms or 10^{-9} g

3.3.6 *s*—seconds

3.3.7 μg —mass in micrograms or 10^{-6} g

4. Summary of Practice

4.1 This practice describes the passive collection and subsequent analysis of soil gas samples, using sorbent samplers to trap VOCs and SVOCs in soil vapor by placing samplers in the subsurface for a period of time at multiple locations across a site. Placement of the sampler can be in open soils (i.e., not covered by a surface such as asphalt or concrete), or advanced through slab surfaces (e.g., parking lots, streets, sidewalks, building slabs, and basement floors) to allow for subslab soil gas sampling. This practice provides standard guidance for passive soil gas (PSG) sampling and analysis performed in support of, but not limited to, site assessment, site monitoring, and vapor intrusion investigations. While several different types and combinations of sorbent materials can be used to trap VOCs and SVOCs in soil gas, this practice is intended to achieve representative and reproducible samples of known quality. The design of PSG surveys (for example, sampler

design, sample spacing, the sampler exposure period, and analytical methods) is within the scope of this practice. These guidelines are not intended to restrict the sampler design or its application in regards to spacing, sampler distribution, or time of exposure; however, these guidelines are meant to provide a general idea of common practice at the time this standard was prepared.

5. Significance and Use

5.1 Passive soil gas samplers are a minimally invasive, easy-to-use technique in the field for identifying VOCs and SVOCs in the vadose zone. Similar to active soil gas and other field screening techniques, the simplicity and low cost of passive samplers enables them to be applied in large numbers, facilitating detailed mapping of contamination across a site, for the purpose of identifying source areas and release locations, focusing subsequent soil and groundwater sampling locations, focusing remediation plans, identifying vapor intrusion pathways, tracking groundwater plumes, and monitoring remediation progress. Data generated from passive soil gas sampling are semi-quantitative and are dependent on numerous factors both within and outside the control of the sampling personnel. Key variables are identified and briefly discussed in the following sections.

NOTE 1—Additional non-mandatory information on these factors or variables are covered in the applicable standards referenced in Section 2, and the footnotes and Bibliography presented herewith.

5.2 *Application*—The techniques described in this practice are suitable for sampling soil gas with sorbent samplers in a wide variety of geological settings for subsequent analysis for VOCs and SVOCs. The techniques also may prove useful for species other than VOCs and SVOCs, such as elemental mercury, with specialized sorbent media and analysis.

5.2.1 *Source Identification and Spatial Variability Assessment*—Passive soil gas sampling can be an effective method to identify contaminant source areas in the vadose zone and delineate the extent of contamination. By collecting samples in a grid with fewer data gaps, the method allows for an increase in data density and, therefore, provides a high-resolution depiction of the nature and extent of contamination across the survey area. By comparing the results, as qualitative or quantitative, from one location to another, the relative distribution and spatial variability of the contaminants in the subsurface can be determined, thereby improving the conceptual site model. Areas of the site reporting non-detects can be removed from further investigation, while subsequent sampling and remediation can be focused in areas determined from the PSG survey to be impacted.

5.2.2 *Monitoring*—Passive soil gas samplers are used to monitor changes in site conditions (e.g., new releases on-site, an increase in contaminant concentrations in groundwater from onsite or off-site sources, and effectiveness of remedial system performance) as reflected by the changes in soil gas results at fixed locations over time. An initial set of data is collected to establish a baseline and subsequent data sets are collected for comparison. The sampling and analytical procedures should remain as near to constant as possible so significant changes in soil gas results can be attributed to those changes in subsurface

contaminant levels at the site that will then warrant further investigation to identify the cause.

5.2.3 Vapor Intrusion Evaluation—Passive soil gas sampling can be used to identify vapor migration and intrusion pathways (see Practice **E2600**), with the data providing a line of evidence on the presence or absence of the compounds in soil vapor, the nature and extent in relation to potential receptors, and whether a vapor pathway is complete. Sorbent samplers can be placed beneath the slab or in close proximity to buildings to collect time-integrated samples targeting VOCs and SVOCs at concentrations often lower than can be achieved with active soil gas sampling methods.

5.3 Limitations—Passive soil gas data are reported in mass of individual compounds or compound groups identified per sample location, with the reporting units generally in nanograms (ng) or micrograms (µg) per sampler and not a concentration (see **6.8**). Ideally, the data produced using this method will be representative of time-weighted soil gas concentrations, present in the vicinity of the PSG sampler and sorbed on the sampler during the exposure period; however, non-uniformity of sampler design, starvation effects during sample collection, or an insufficient amount of sorbent that results in saturation of the sorbent surface area, or combinations thereof, will affect the relationship between sorbed mass and soil gas concentrations present. The degree to which these data are representative of any larger areas or different times depends on numerous site-specific factors. In general, information obtained from a passive soil gas sampling program alone is not sufficient to support a quantitative determination of soil gas concentrations.

5.4 Sampler Design—Passive soil gas is an effective investigatory/monitoring tool if the appropriate quality controls are included in the technology design, which includes uniformity in the construction of the sampler. At a minimum, controls should be in place to ensure that (1) the appropriate sorbents with hydrophobic properties are used to target the compounds of concern (see Practice **D6196**), (2) materials used to house the sorbents are chemically-inert, non-reactive or corrosive, and will not off-gas compounds or act as competing sorbents (see Guide **D5314**, paragraph 6.5.3), and (3) the sorbents are housed in suitable containers that protect the sorbents, allow diffusion of the soil gas to the sorbents, and facilitate installation of the sampler to the desired sampling depth.

5.4.1 Sampler Conditioning—Before being sent to the field for deployment, the PSG sampler should be conditioned to remove any potential contamination present on or in the sorbent and sampler materials or both encountered during sampler construction or storage prior to use. The conditioning process should be one that does not damage the sorptive capability of the sorbent. Following conditioning, the sampler is then capped/resealed and stored in a container that provides adequate protection against ambient sources of contamination before and after sample collection in the field, including during transport. Preparation blanks from each batch of conditioned samplers should be analyzed to verify that the sorbents were effectively conditioned and do not retain measurable masses of target compounds above reporting limits. Furthermore, when trip blanks, which are included with all shipments to and from

the field, report non-detects for the targeted compounds, these QC samples provide additional evidence that the samplers were conditioned to have no measurable mass of target compounds and that the measurements on field samples originate from the site itself.

5.5 Sampler Exposure Periods—Guidelines for PSG exposure periods for source identification, spatial variability assessment, and vapor intrusion evaluation should consider the project objectives, target compounds, required detection limits or anticipated soil gas concentrations or both, design of the passive sampler, matrix heterogeneity, soil types (total porosity), soil moisture level (water filled porosity), and depth to expected contaminants. Sites having coarse-grained dry soils, high concentrations, shallow groundwater or soil contamination or both, and volatile compounds typically require shorter exposure periods. Sites with finegrained, clays or moist soils or both, deep contaminant sources, low concentrations, or SVOCs, or combinations thereof, typically require longer exposure periods. Exposure periods typically range from days to weeks but can be as brief as one hour when high concentrations of target compounds are expected in the soil vapor.

5.6 Sampler Spacing—Grid designs can consist of regularly spaced sampler locations, random or irregular spaced, and as transects or varying spatial intervals (see Guide **D6311**). Biased spacing in which smaller sample spacing is used in areas with known or suspected targets (that is, source areas) and large spacing in areas not believed to be impacted are also used. For large area investigations, a staged or phased sampling program can be used. The investigation begins with a widely spaced regular grid design. The initial soil gas results are reviewed and subsequent sampling is conducted at locations where the target compounds were observed. The subsequent survey design consists of more closely spaced samples to resolve the feature of interest in greater detail. Multiple phases of soil gas sampling can be combined to provide one comprehensive image of the soil gas results. Staged or phased investigations require multiple deployments adding costs to the overall investigations. However, areas of the site that have nondetectable values in the soil gas may be removed from further investigation.

5.6.1 There is no prescribed or set sampler spacing appropriate for all sites, as sample spacing and survey design are based on project objectives and each site is unique. General recommendations for sampler spacing range from 3 to 30 m, with 7.5- to 15-m spacing when site knowledge is lacking. Infill sampling is recommended in areas having wider sample spacing initially.

5.6.2 Site-specific information (investigation area size, groundwater depth, soil type and moisture content, purpose of the investigation, *etc.*) should be considered along with these guidelines in determining the grid spacing used. The selection of grid cell size (a direct function of the sampler spacing deployed in a grid pattern) is strongly dependent upon the relationship between both project confidence level and budget requirements. The tendency exists for investigators with constrained budgets to use overly large grid cell spacing. This action of “undersampling” normally results in inadequate, over-interpreted data with unsupported conclusions. Care shall

be taken to avoid this problem (Guide [D5314](#)). In designing an effective soil gas survey to develop a rational conceptual site model, the survey objective balanced by budget should determine the sample spacing.

5.7 Sampling Depth—Consideration of project objectives should be taken into account when determining deployment depth. It is ideal, when possible, to deploy samplers at the same depth to ensure data consistency. PSG samplers are generally installed from a depth of 15-cm to 1.0-m BLS; however, holes may be advanced to greater depths when appropriate, and samplers can also be suspended beneath surface flux chambers or in permanent vapor ports.

5.8 Soil Types—In general, sandy soils tend to be more porous and permeable and, thus, require shorter exposure times. Conversely, soils with high clay contents tend to be less porous and permeable and typically have lower flux rates (see Practice [D2487](#)). Soil types vary in vapor permeability due to the differences in the number and interconnectivity of air-filled pores. The more air-filled, interconnected the pores are, the greater the potential flux of contaminants through the soil to the sampler. Starvation effects resulting in low bias are more likely to occur in low permeability soils where the flux through the soil matrix is limited.

5.9 Effects of Soil Moisture—Because diffusion of vapors from subsurface sources to passive samplers relies on interconnected and air-filled pores within the soil column, soil moisture can have a significant effect on the flux of contaminants and, therefore, the mass of the contaminant available for adsorption by the sampling device. The use of hydrophobic sorbents minimizes the effect on sampler sensitivity, but does not change the impact of soil moisture on contaminant soil gas concentrations. As a result, areas of high soil moisture may have significantly lower soil gas results than areas of low soil moisture, even though subsurface concentrations are similar in both areas. Therefore, some knowledge of the soil moisture conditions can help in interpreting soil gas results. This knowledge is also useful for comparing results from subsequent surveys performed at a site.

5.10 Effects of Target Compounds—In general, the larger the molecular weight of the compounds being targeted, the lower the vapor pressure and resulting concentrations in the soil gas, and therefore, the longer the required exposure time of the PSG samplers in the vadose zone.

5.11 Sealing (Plugging) the Top of the Hole—Once the PSG sampler is inserted in the ground, the top of the hole is plugged with a material that will effectively seal the hole, such as aluminum foil or cork, which can then be covered with soil. For concrete or asphalt surfacing, an approximately 5-mm-thick mortar or quick-setting concrete patch above the plug can be used as an option to maintain the integrity of the surface while the sampler is in the ground. The materials used to plug the hole should not contribute compounds of concern and the seal should be flush mounted to keep the sampler safe from harm, prevent ingress of ambient air or surface water, and not interrupt ongoing site activities during the exposure period.

5.12 Effects of Ambient Air While Installing/Retrieving Samplers—PSG samplers arrive at the site sealed to protect the

sorbents from contaminants in ambient air during transport. Just prior to installation into the hole, and then again during retrieval, the sampler is exposed to ambient air for a brief period of time. The typical time of exposure to the ambient air is less than 15 s. In some instances, it may be necessary to collect a field blank using a PSG sampler to evaluate whether compounds in the ambient air potentially biased the results. To perform this quality control check, an identical PSG sampler is opened and exposed to the ambient air for approximately the same amount of time required to install and then later retrieve a PSG sampler at a designated location. The field blank is sealed at all other times and is transported to the laboratory along with the field samples. Care should be taken to minimize the sorbent exposure to ambient air during field activities. Obvious sources of contamination (e.g., gas-powered electrical generators or vehicle exhaust) should not be in close proximity when installing/retrieving a sampler.

NOTE 2—The quality of the result produced by this standard is dependent on the competence of the personnel performing it and the suitability of the equipment and facilities used. Agencies that meet the criteria of Practice [D3740](#) are generally considered capable of competent and objective testing/sampling/inspection/and so forth. Users of this standard are cautioned that compliance with Practice [D3740](#) does not in itself assure reliable results. Reliable results depend on many factors; Practice [D3740](#) provides a means of evaluating some of those factors.

6. Procedure

6.1 PSG Sample Collection—The overall passive soil gas survey design, including the sampling depth and the time of exposure in the field, should be determined based on the guidelines given in Section [5](#) and discussions with the service provider, environmental consultant, state or federal authorities, and other stakeholders.

6.1.1 Typically, a small-diameter (approximately 2.5 cm) hole that is approximately 15 cm to 1.0 m deep is advanced at each of the sample locations. Holes are advanced using a hammer drill with drill bits, a slide hammer, hand auger, direct push technology, or other comparable equipment. For locations covered with asphalt or concrete, the hole is drilled through the surfacing to the underlying soils. Depths greater than 1.0 m may require more invasive drilling techniques, as well as possibly casing the hole to keep it open to PSG sampler insertion

6.1.2 The sorbent sampler is then lowered into the hole to the desired sampling depth using wire or string, which facilitates retrieval of the sampler following the exposure period.

6.1.3 The upper portion of the hole is plugged, as described in [5.11](#), to protect the sorbent sampler from the ingress of ambient air or surface water and not interrupt ongoing site activities during the exposure period.

6.1.4 Field notes are recorded during the installation and retrieval of the samplers. Installation and retrieval logs can take a variety of forms. At a minimum, field notes should include sampler ID, deployment and retrieval dates and times, a site map with sampler location or GPS coordinates, and any other relevant site information.

6.2 Sampler Exposure Period—The length of time that the samplers remain in the ground to adsorb compounds in soil gas should be determined before field mobilization using the guidelines given in [5.5](#). Namely, consideration shall be given to

the project objectives, target compounds, required detection limits or anticipated soil gas concentrations or both, soil types (total porosity), soil moisture level (water-filled porosity), and depth to expected contaminants.

6.3 PSG Sampler Retrieval—Following the exposure period, the samplers are retrieved from the sample holes, capped/resealed, labeled, packaged, and returned to the laboratory for analysis with accompanying chain-of-custody procedures. The chain-of-custody should record project and client information (e.g., site reference, client contact, and sample identification numbers), and include signature sections. The chain-of-custody can be organized to include an installation log or field deployment report, which would include more sampler specific information (e.g., time and date of sampler deployment and retrieval, sample location conditions), or these can be separate documents. Examples of a field deployment report and a chain-of-custody form are shown in **Figs. 1 and 2**. Materials used to label samplers should not emit obvious odors, such as permanent markers, which contain solvents; however, ballpoint pens are acceptable.

6.4 Shipment of Samplers—The use of preservatives and refrigerants is not necessary and is not recommended during shipment of PSG samplers because of the potential to introduce contaminants into the shipping container. When packing the samplers for shipment, cushioning materials that may off-gas compounds also should not be used (e.g., styrofoam packaging or newsprint). Samplers should be shipped to the laboratory using expedited courier service with overnight delivery being the preferred method. Once at the laboratory, when stored in a clean, safe area, most sorbent samplers have a relatively long holding time. However, holding times vary depending on sorbents used, sampler seal/cap, storage container, and method of storage.

6.5 Field Quality Control (QC) Samples—There are three field QC samples that can be incorporated into a PSG survey: duplicates, trip blanks, and field blanks.

6.5.1 Field sample duplicates provide an assessment of variability associated with field sampling, sampler performance, and the analytical method. A field sample duplicate may be obtained by co-locating PSG samplers (i.e., collecting samplers in separate holes in close proximity to one another) or by using replicate sorbents housed side-by-side within the sampler. Analysis of replicate sorbents within the sampler provides information as to the variability of the sampler performance and the analytical method. Analysis of co-located samplers additionally provides information on variability of the soil matrix. The percentage of duplicate field samples is determined by the data quality objectives.

6.5.2 Trip blanks are PSG samplers that are prepared, transported, and analyzed with other samples but intentionally not exposed. These QC samples can be used to identify background contamination contributed during the transport and storage of the samplers. It is recommended that one trip blank, at a minimum, be included with each batch of PSG samplers during shipment to the field and when submitted to the laboratory.

6.5.3 A field blank is a PSG sampler that is exposed to ambient air at a preselected location during both installation

and retrieval of samplers for the approximate amount of time that a sampler is exposed to ambient air while being installed and retrieved. These QC samples are typically not required unless high concentrations of targeted compounds are anticipated to be present in the ambient air during field activities (e.g., sampling inside an operating manufacturing facility where solvents are being used).

6.6 Decontamination of Equipment—Field equipment used to create sampling holes and insert samplers, as required, should be decontaminated between sample locations when contaminated soil at the near surface is anticipated to avoid cross-contamination (see Practice **D5088**).

6.7 Analysis—Several options for the analysis of PSG samples are described in Guide **D5314** to meet project data quality objectives and budgets. Thermal desorption and solvent extraction are the two most common methods for introduction of the sample as part of the analytical method. The laboratory should minimize the volume of stored solvents present in the building where analysis takes place and implement controls to maintain indoor air quality that will not bias the samples. Analysis of the sorbent housed in the PSG sampler should be in accordance with established methods that are approved before initiation of the project. The selected method must be documented, defensible, produce data of known quality, and meet the data quality objectives of the investigation (see Practice **D5792**). Typical analytical methods include, but are not limited to, Practice **D6196**, U.S. EPA Method 8260C, Method 8270C, and Method TO-17; however, other sample extraction and analytical methods may be performed if the methods meet the project's data quality objectives. The level of QA/QC specified by the referenced method should be included to produce a valid, defensible dataset. Regardless of the analytical method utilized, deviations from the referenced method shall be noted and qualified in the reported PSG analytical data package.

6.8 Data Usage—Passive soil gas data are reported in mass of individual compounds or compound groups identified per sample location, with the reporting units generally in nanograms (ng) or micrograms (μg) per sampler. Results from a passive soil gas survey typically are then used to guide where follow-on samples should be collected to obtain corresponding concentrations of the contaminants in soil, soil gas, or groundwater, or combinations thereof, as well as eliminate those areas where more intrusive samples are not required. These correspondent values furnish the basis for approximating a relationship between the passive soil gas data and the subsequently collected data, which can then be extrapolated across the entire survey grid to estimate concentrations at other locations. However, when comparing PSG data to soil or groundwater data, specific conditions at individual sample points, including, but not limited to, soil porosity and permeability, depth to contamination, and perched groundwater, can influence soil gas measurements at individual locations. In addition, the compound mass from the passive sampler may originate from free-phase volatilization, off-gassing from soil adsorbed contamination, or partitioning of compounds out of solution and into the vadose zone from the contaminated groundwater, or combinations thereof. Each or